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## Marine fungal diversity: a comparison of natural and created salt marshes of the north-central Gulf of Mexico

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**Abstract:** Marine fungal communities of created salt marshes of differing ages were compared with those of two reference natural salt marshes. Marine fungi occurring on the lower 30 cm of salt marsh plants *Spartina alterniflora* and *Juncus roemerianus* were inventoried with morphological and molecular methods (ITS T-RFLP analysis) to determine fungal species richness, relative frequency of occurrence and ascospore density. The resulting profiles revealed similar fungal communities in natural salt marshes and created salt marshes 3 y old and older with a 1.5 y old created marsh showing less fungal colonization. A 26 y old created salt marsh consistently exhibited the highest fungal species richness. Ascospore density of the dominant fungal species on each host was significantly higher in natural marshes than in created marshes at all three sampling dates. This study indicates marine fungal saprotroph communities are present in these manmade coastal salt marshes as early as 1 y after marsh creation. The lower regions of both plant hosts were dominated by a small number of marine ascomycete species consistent with those species previously reported from salt marshes of the East Coast of USA.

**Key words:** ITS T-RFLP analysis, *Juncus roemerianus*, salt marsh fungi, *Spartina alterniflora*

### INTRODUCTION

Ascomycetous fungi are the principle decomposers of the emergent salt marsh plant *Spartina alterniflora* (smooth cordgrass; Poaceae) (Newell 1996) and also are prevalent in the decay of *Juncus roemerianus* (black needlerush; Juncaceae) (Kohlmeyer and Volkman-Kohlmeyer 2001). Together these two plant species represent the dominant Gulf Coast salt marsh vegetation (Eleuterius 1981). Lignocellulose comprises 70–75% of the organic mass of dry shoots of *S. alterniflora* (Newell and Porter 2000). Other trophic levels depend on the extracellular enzymes of saprotrophic fungi to cleave lignocellulose into

smaller molecules that can enter the food web via mycophagic invertebrates and bacteria (Newell and Porter 2000).

Salt marshes provide critical habitat for a variety of species but are currently in decline worldwide due to a variety of natural and anthropogenic causes. Restoration efforts are under way along the north-central Gulf Coast with nursery-raised vegetation. This study was part of a larger multitrophic search to identify assessment metrics suitable for measuring marsh restoration success. Created tidal marshes have been shown to lack organic matter and nutrients when compared with natural salt marshes (Zedler and Lindig-Cisneros 2000), so it is imperative to consider the often underappreciated role of saprotrophic fungi when restoring these systems. As the major contributor to denitrification and mineralization of carbon and nitrogen (Lillebo et al. 1999) fungi are vital to salt marsh nutrient cycling processes.

*Spartina alterniflora* and *J. roemerianus* do not abscise their leaves, and decay occurs while the plants are still standing (Newell and Palm 1998). Saprotrophic fungal species are well characterized for natural salt marshes of the eastern USA (Kohlmeyer and Kohlmeyer 1979, Kohlmeyer and Volkman-Kohlmeyer 1991, Newell and Porter 2000). However no comparison of fungal communities has been made between natural and created salt marshes or between East Coast and Gulf Coast salt marshes of USA.

The objectives of this study were to compare marine fungal saprotroph communities among natural salt marshes and created salt marshes of different ages to determine when the marine fungal community establishes in created salt marshes and to determine whether fungal community profiles can serve as indicators of restoration success. This study also sought to provide valuable geographic distribution data for marine fungi along the understudied north-central Gulf Coast and compare the species found with those known from salt marshes of the East Coast. Effects of abiotic factors and host plant on fungal assemblages also were explored.

### MATERIALS AND METHODS

*Collection and morphological techniques.*—Saprotrophic marine fungal species associated with standing salt marsh vegetation were inventoried with morphological and molecular techniques (ITS T-RFLP analysis) to determine ascospore density, species richness and relative frequency of

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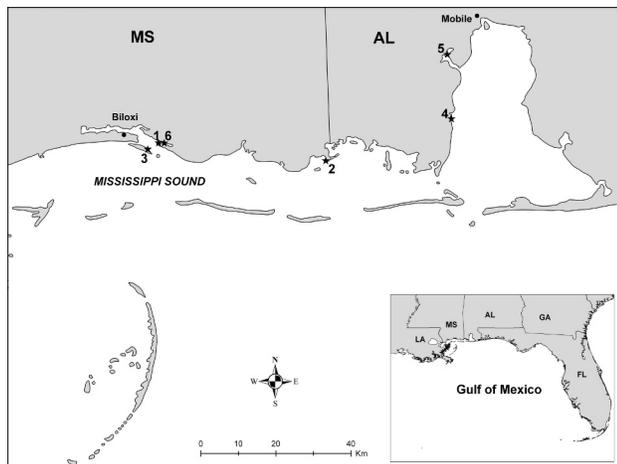


FIG. 1. Location of salt marsh sampling sites along the north-central coast of the Gulf of Mexico and created marsh age at time of collection. Natural sites: site 1 = Shack Island, Mississippi; site 2 = Grand Bay National Estuarine Research Reserve, Mississippi; site 3 = Deer Island, Mississippi, 1 y; site 4 = Mon Luis Island, Alabama, 1.5 y; site 5 = Dog River, Alabama, 3 y; site 6 = Davis Bayou, Mississippi, 26 y.

occurrence over three sampling seasons, fall 2005, spring 2006 and fall 2006. Because fungal species are known to exhibit a distinct vertical zonation on salt marsh plants based on halotolerance (Gessner 1977) only the lower 30 cm of plants was inventoried for marine fungi.

Decaying leaves of *Spartina alterniflora* and *Juncus roemerianus* were collected from two natural salt marshes and four created salt marshes along the north-central Gulf Coast (FIG. 1). Three GPS points (sites) were marked at each marsh and water temperature (C), pH, salinity (ppt) and dissolved oxygen (%) were recorded at each site. Twelve decaying leaf blades from each plant species were collected at random in a 5 m radius from each GPS point, for a total of 36 leaves of each plant collected from each marsh. Leaves were placed in new resealable polyethylene bags and transported on ice to the laboratory. In the laboratory each leaf blade was cut in half with one-half used for morphology (ascomata densities, fungal identifications and isolations) and one-half used for ITS T-RFLP analysis.

Morphological identification and quantification of ascomata were carried out with light microscopy to directly observe fungal reproductive structures on decaying *S. alterniflora* and *J. roemerianus*. Ascomata were removed from plant leaves with the aid of a flame-sterilized needle and squash mounted in sterile distilled water or lactophenol cotton blue and examined with a Nikon Eclipse 80 microscope with Nomarski interference contrast optics. Indirect observation with the spore-trapping method of Newell (2001) also was performed. Digital photographs of microscopic fungal structures were taken with a SPOT Insight camera (Diagnostic Instruments Inc., Sterling Heights, Michigan), and measurements were made with SPOT 4.1 software (Diagnostic Instruments Inc., Sterling Heights, Michigan). Number of ascomata per mm<sup>2</sup> leaf area

of the dominant fungal species was recorded for 12 6 cm leaf blade segments from each site.

**DNA extraction and sequencing.**—An ascospore expulsion technique modified from Newell (2001) was used to obtain fungal single-spore isolates from leaves of both plants. Ascospores of ascomycetes that did not forcibly eject spores were isolated by nichrome-needle micromanipulation of ascotal contents and dilution plating onto antibiotic saltwater agar (ASWA) plates. Fungal cultures were maintained in the dark at 4 C. Culture slants were maintained on saltwater potato dextrose agar (SWPDA) at 4 C.

DNA was extracted from single-spore isolates with a DNeasy Plant Mini Kit (QIAGEN, Germantown, Maryland) with the addition of lyticase (Sigma, St Louis, Missouri) following the method of Raja et al. (2003). DNA was amplified with fungal ITS primers ITS 1-F (Gardes and Bruns 1993) and ITS 4-A (Larena et al. 1999) (Invitrogen, San Diego, California). This primer pair has enhanced specificity for members of phylum Ascomycota (Larena et al. 1999), which are known to dominate the decay of standing salt marsh vegetation (Newell and Porter 2000). These primers amplified a total product of 1225 bp which included the complete ITS region (ITS 1 and 2, and 5.8S [583 bp]), as well as the 3' end of the 18S and the 5' end of the 28S rDNA. PCR was performed with these cycling parameters: a 3 min initial denaturation step at 95 C, followed by 35 cycles of 1 min at 95 C, 30 s annealing at 52 C and 1 min extension at 72 C, ending with a 10 min final extension at 72 C in a Px2 thermal cycler (Thermo Fisher Scientific Inc., Waltham, Massachusetts). PCR products were cleaned with a QIAquick PCR purification kit (QIAGEN, Germantown, Maryland). All samples were sequenced at the University of Illinois Urbana-Champaign Core Sequencing Facility with an Applied Biosystems 3730xl DNA Analyzer. Raw sequences were edited, contiguous DNA segments were assembled and consensus sequences were exported with Sequencher 4.7 (Gene Codes Corp., Ann Arbor, Michigan). Sequences were compared with the reference sequence database NCBI GenBank with the BLAST engine (Altschul et al. 1997).

**T-RFLP.**—The terminal-restriction fragment length polymorphism (T-RFLP) technique was used to create site-specific fingerprints. The ITS region of ribosomal DNA was chosen due to its highly conserved nature, combined with its high sequence variability at the species level (Larena et al. 1999). Environmental DNA was extracted from 6 cm leaf blades with an UltraClean™ Soil DNA Isolation Kit (MoBio Laboratories, Carlsbad, California). Fungal ITS rDNA was amplified with the fungal-specific primers ITS 1-F and ITS 4-A with ITS 1-F labeled on the 5' end with the fluorescent dye FAM (6-carboxyfluorescein) (Invitrogen). Target specificity was determined by agarose gel electrophoresis and reaction yield was quantified with a NanoDrop ND-1000 spectrophotometer. PCR products were cleaned with a QIAquick PCR purification kit (QIAGEN, Germantown, Maryland), followed by restriction digestion with the *Hae*III restriction enzyme (Roche Applied Science, Indianapolis, Indiana). For each sample a restriction digest reaction was set up in a 1.5 mL microcentrifuge tube containing either

10 ng (isolate) or 100 ng (community) purified PCR product, 1  $\mu$ L SuRE/Cut buffer M (10 $\times$ ) (Roche), 1  $\mu$ L (10 U) *Hae*III restriction enzyme (Roche) and ddH<sub>2</sub>O to equal a 10  $\mu$ L total reaction volume. Restriction digest reactions were incubated at 37 C for 3 h with a dry bath incubator. After digestion PCR products were stored at -20 C until analysis. All restriction digest reagents were stored at -20 C.

The recognition site for *Hae*III is the DNA nucleotide sequence GGCC; the restriction enzyme cleaves between the second and third nucleotide (G/C), generating multiple restriction fragments for each fungal species. Only the terminal restriction fragment (T-RF) is fluorescently labeled and its size is detected by capillary electrophoresis. Buchan et al. (2002) indicated the restriction enzyme *Hae*III best discriminated between salt marsh ascomycete species because the size of the T-RF produced is conserved within species yet is sufficiently variable to discriminate among species. Thus each T-RF corresponds to one fungal species and therefore the number of T-RFs in a fungal community profile can be used as a proxy for species richness.

Fragment samples were analyzed at the University of Illinois Urbana-Champaign Core Sequencing Facility with an Applied Biosystems 3730xl DNA Analyzer. Each fluorescently labeled PCR reaction was done in triplicate, and the resulting three chromatograms were overlaid for each sample with GeneMapper™ 3.7 software (Applied Biosystems Inc., Foster City, California). Peaks were standardized with the Microsoft Excel-based Macro program TRFLP Tools (Stepanauskas et al. 2002). Peak height in T-RFLP chromatograms was used as a proxy for the relative abundance of fungal taxa represented by restriction fragments (Stepanauskas et al. 2003). Peaks were assumed to be artifacts and removed from analysis if they did not contribute more than 1% to the sum of peak heights in any individual profile and occurred in less than three profiles (Stepanauskas et al. 2003). In addition peaks smaller than 50 bp or larger than 550 bp were assumed to be primer and uncut ITS sequences, respectively, and therefore were discarded from the analyses. We compiled a list of fungal isolates obtained in this study and their T-RF sizes (TABLE II). In all 60 T-RFLP chromatograms were generated (33 from *S. alterniflora*, 27 from *J. roemerianus*). Three chromatograms were generated for each site and overlaid, resulting in 20 overlays (11 *S. alterniflora*, 9 *J. roemerianus*), corresponding to one overlay per plant per site per collection (fall 2005, spring 2006 and fall 2006). T-RF species richness was calculated from each overlay with significant peaks as proxies for individual ascomycete species.

Statistical analyses were performed with SPSS 12 (SPSS Inc., Chicago, Illinois) with results considered significant if  $P < 0.05$ . Datasets were tested for homogeneity of variance and normality with Levene's test for equality of variance and the Kolmogorov-Smirnov one-sample test for normality respectively. The effect of site age on fungal species richness was evaluated with linear regression. Between-site comparisons of the number of ascomata of the dominant marine saprotrophic fungal species per mm<sup>2</sup> leaf blade were made with one-way ANOVA. Seasonal differences in ascomata

density were evaluated with independent samples *t*-tests. The effect of season on fungal species richness was investigated with one-way ANOVA.

## RESULTS

Six species of marine fungi were detected and identified morphologically on decaying *Spartina alterniflora* (TABLE I). *Phaeosphaeria spartanicola* was found in all collections from both natural and created salt marshes. *Buergenerula spartinae* (FIG. 2C) was the second most commonly encountered species, present in all collections from natural marshes and only absent from two of the created sites, Davis Bayou Island, Mississippi, (site 6) in fall 2005, and Deer Island, Mississippi, (site 3) in fall 2006. *Phaeosphaeria halima* was only detected in fall 2005 at one natural site, Shack Island (site 1), and one created site, Davis Bayou (site 6). Ascospores of this species were 16–17  $\times$  5  $\mu$ m (FIG. 2A). *Leptosphaeria pelagica* was encountered only at one created site, Davis Bayou Island (site 2), in fall 2005.

Several of the fungal species detected in this study deviated from published descriptions or were new records for the region. *Mycosphaerella* sp. I and II have been documented from East Coast salt marshes for more than 28 y but have not been described formally (Kohlmeyer and Kohlmeyer 1979, Buchan et al. 2002). In the present study *Mycosphaerella* sp. I was detected in three collections from two natural marshes, Shack Island (site 1) in fall 2005 and Grand Bay NERR (site 2) in spring and fall 2006. This species was not detected morphologically at any of the created marshes. *Mycosphaerella* sp. I was described as having hyaline ascospores (Kohlmeyer and Kohlmeyer 1979), while in this study some were hyaline and some became yellow-brown with age (FIG. 2B, E), commonly noted for genus *Mycosphaerella* (Kohlmeyer and Kohlmeyer 1979) but not this species. *Mycosphaerella* sp. II was detected in five collections: from the natural site, Shack Island (site 1), in fall 2005 and spring 2006; the created sites, Deer Island (site 3), in spring 2006, Mon Luis Island (site 4), and Davis Bayou (site 6), in fall 2006 (FIGS. 1, 2D). Ascomata of *Mycosphaerella* sp. II were consistently encountered near ascomata of the dominant saprotrophic ascomycete *P. spartanicola*. Here we report *Mycosphaerella* sp. I and II from Gulf Coast salt marshes for the first time.

Five species of marine fungi were identified morphologically on *Juncus roemerianus* (TABLE I). *Massarina ricifera* was the most common marine fungal saprotroph detected on *J. roemerianus* and was present in all collections. Asci were bitunicate, short pedunculate, eight-spored and 71–72  $\times$  14–15  $\mu$ m

TABLE I. Fungal species by plant host, site and season, with relative frequency of occurrence detected morphologically

	Fall 2005		Spring 2006				Fall 2006						Relative frequency at natural sites (%)	Relative frequency at created sites (%)	Relative frequency in all collections (%)	
	Site:	1	6	1	2	3	6	1	2	3	4	5				6
<i>Spartina alterniflora</i> (n = 60 for natural sites, n = 72 for created sites)																
<i>Phaeosphaeria spartinicola</i>																
Leuchtman	+	+	+	+	+	+	+	+	+	+	+	+	+	100	100	100
<i>Buergenerula spartinae</i> Kohlm. & Gessner																
	+		+		+		+	+	+				+	47	35	82
<i>Mycosphaerella</i> sp. II	+		+		+			+		+				73	40	55
<i>Mycosphaerella</i> sp. I	+			+	+			+						17	10	27
<i>Phaeosphaeria halima</i> (T.W. Johnson) Shoemaker & C.E. Babc.																
	+	+												20	16	18
<i>Leptosphaeria pelagica</i> Jones			+											0	17	9
<i>Juncus roemerianus</i> (n = 48 for natural sites, n = 60 for created sites)																
<i>Massarina ricifera</i> Kohlm., Volkm.-Kohlm. & O.E. Erikss.																
			+	+	+	+	+	+		+	+	+		100	100	100
<i>Phaeosphaeria roemeriani</i> Kohlm., Volkm.-Kohlm. & O.E. Erikss.																
								+						2	0	1
<i>Anthostomella poecila</i> Kohlm., Volkm.-Kohlm. & O.E. Erikss.																
												+		0	2	1
<i>Mycosphaerella</i> sp.	+													2	0	1
<i>Leptosphaeria</i> sp.	+													2	0	1

(FIG. 2F). Ascospores were three-septate, strongly constricted at septa, hyaline to light yellow-brown, had one guttule in each section and were  $21\text{--}22 \times 6 \mu\text{m}$  (FIG. 2G). *Massarina ricifera* is also common in East Coast salt marshes where it has been collected every month of the year (Kohlmeyer et al. 1995). However in the present study no gelatinous sheath was observed when ascospores were examined in India ink, contrary to the two-layered gelatinous sheath noted in the original description (Kohlmeyer et al. 1995).

*Anthostomella poecila* was detected only at one created site, Mon Luis Island (site 4), in fall 2006. Ascospores were brown to dark brown, broadly ellipsoidal and  $14\text{--}17 \times 7\text{--}9 \mu\text{m}$  (FIG. 2L). A species of *Mycosphaerella* was found at one of the natural sites, Shack Island (site 1), in fall 2006 and was identified to genus by its solitary, erumpent, ostiolate globose ascumata, absence of paraphyses, eight-spored elongate-cylindrical bitunicate asci and brown, one-septate  $10\text{--}14 \times 5\text{--}6 \mu\text{m}$  ascospores (Kohlmeyer and Kohlmeyer 1979, Kohlmeyer and Volkmann-Kohlmeyer 1991, Hyde and Sarma 2000). A BLAST analysis of the ITS rDNA of a single-spore isolate sequenced in the current study revealed a 98% match with a *Mycosphaerellaceae* species isolated from Hawaiian seawater.

*Leptosphaeria* sp. was detected in one collection at one natural site, Shack Island (site 1), in spring 2006.

Ascumata were large, black, gregarious and emergent. Asci were  $58\text{--}60 \times 10\text{--}11 \mu\text{m}$  and were bitunicate and eight-spored (FIG. 2H). Ascospores were hyaline, three-septate,  $16\text{--}18 \mu\text{m} \times 4 \mu\text{m}$  and surrounded by a gelatinous sheath (FIG. 2I). *Phaeosphaeria roemeriani* was found only at one natural site, Shack Island (site 1), in fall 2006. It matched the published description (Kohlmeyer et al. 1998) except that ascospore guttules were absent (FIG. 2J, K).

The present study detected a *Mycosphaerella* species occupying the lower 30 cm of one *J. roemerianus* plant collected from the created site, Deer Island, Mississippi, (site 3) in spring 2006. Ascospores were on average smaller than spores of *Mycosphaerella* sp. II collected on *S. alterniflora* ( $10\text{--}14 \times 5\text{--}6 \mu\text{m}$  compared with  $15\text{--}20 \times 5\text{--}8 \mu\text{m}$ ) and were dark brown, not hyaline. The only species of *Mycosphaerella* reported from *J. roemerianus* is *M. euryptami*, described from the upper senescing leaves and not reported from the tidally inundated region (Kohlmeyer et al. 1999). The *Mycosphaerella* species reported in this study does not fit the published description of *M. euryptami*, which has ascospores  $23\text{--}29 \times 5.5\text{--}6.5 \mu\text{m}$ ; it also does not fit the published descriptions of the four other *Mycosphaerella* species known from the marine environment (Kohlmeyer and Kohlmeyer 1979) and therefore might be a new marine species.

A species of *Leptosphaeria* collected on *J. roemer-*

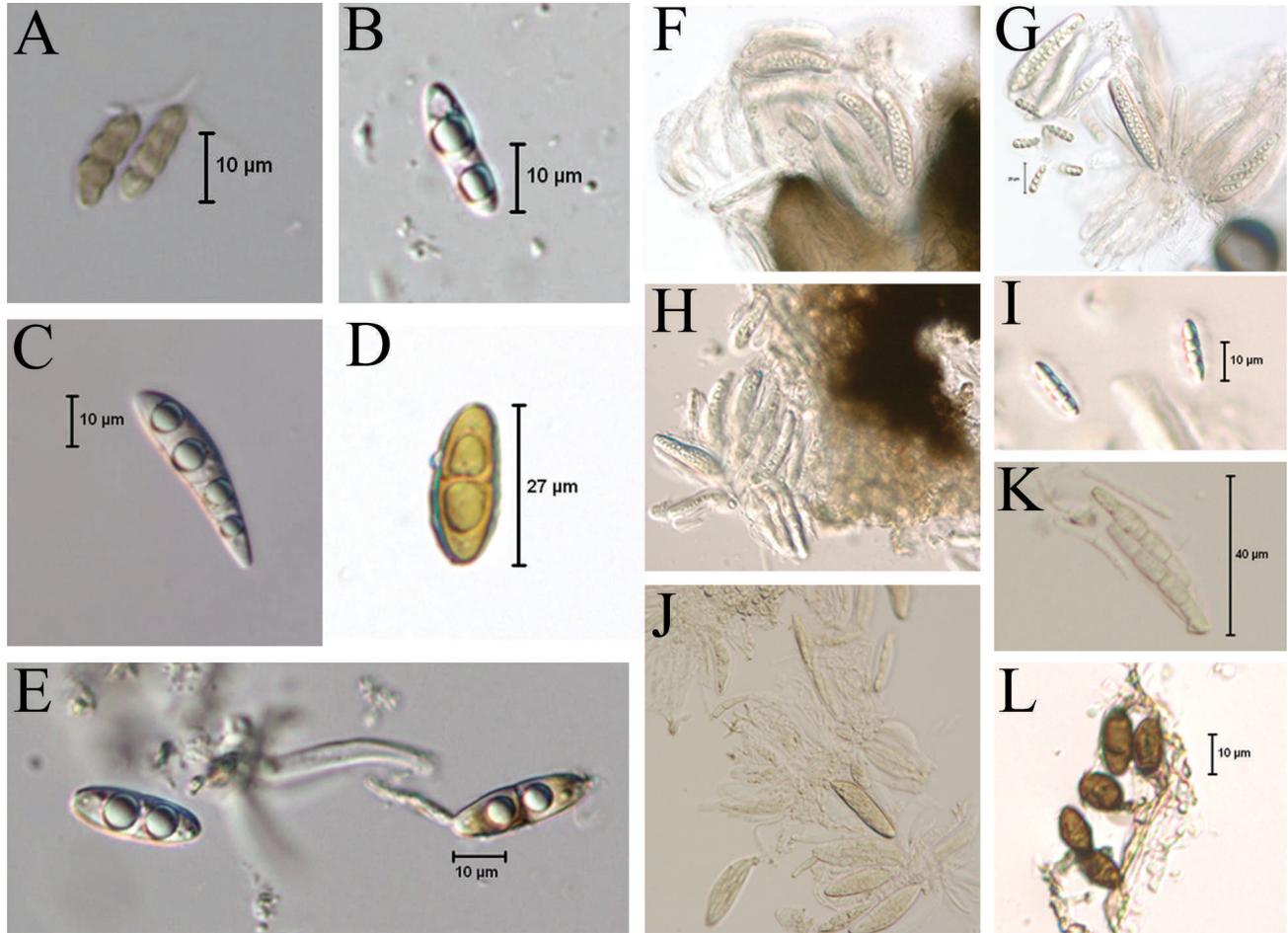


FIG. 2. A–E. Saprotrrophic marine fungi on *Spartina alterniflora*. A. Ascospores of *Phaeosphaeria halima*. B. Ascospore of *Mycosphaerella* species I. C. Ascospore of *Buergenerula spartinae*. D. Ascospore of *Mycosphaerella* species II. E. Ascospores of *Mycosphaerella* sp. I. Young hyaline ascospore at left, on right older germinating ascospore is turning brown. F–L. Saprotrrophic marine fungi on *Juncus roemerianus*. F. Ascoma and asci with ascospores of *Massarina ricifera*, the dominant saprotroph on *J. roemerianus*. G. Asci and ascospores of *Massarina ricifera*. H. Squash mount showing asci and ascospores of *Leptosphaeria* sp. I. I. Ascospores of *Leptosphaeria* sp. J. Squash mount of ascoma of *Phaeosphaeria roemeriani*. K. Ascospores of *Phaeosphaeria roemeriani*. L. Ascospores of *Anthostomella poecila*.

*ianus* in spring 2006 had hyaline  $58\text{--}74 \times 10\text{--}11 \mu\text{m}$  bitunicate asci (FIG. 2H) and hyaline  $16\text{--}18 \times 4 \mu\text{m}$  ascospores (FIG. 2I). *Leptosphaeria marina* has been reported from this host; however its asci and ascospores are respectively  $112\text{--}220 \times 18\text{--}35 \mu\text{m}$  and  $35\text{--}68 \times 8\text{--}12 \mu\text{m}$  (Kohlmeyer and Kohlmeyer 1979). The *Leptosphaeria* species detected on *J. roemerianus* in this study does not match any published description and might be a new species.

Significantly more ascomata of the dominant marine fungus *Phaeosphaeria spartinicola* were found on leaves of *S. alterniflora* from natural salt marshes ( $4.45 \text{ ascomata}/\text{mm}^2$ ) than from created marshes of all ages ( $0.39 \text{ ascomata}/\text{mm}^2$ ) (ANOVA:  $F = 46.780$ ,  $P = 0.000$ ) ( $n = 132$ ). Similarly significantly more ascomata of the dominant *J. roemerianus* saprotroph, *Massarina ricifera*, were found in natural marshes

( $7.37 \text{ ascomata}/\text{mm}^2$ ) than in created marshes ( $1.48 \text{ ascomata}/\text{mm}^2$ ) (ANOVA:  $F = 955.367$ ,  $P = 0.000$ ) ( $n = 108$ ). However the mean number of ascomata produced in spring 2006 was not significantly different from the mean number of ascomata in the fall 2006 collection [ $t(107) = 0.390$ ,  $P = 0.698$ ].

We calculated T-RF species richness (FIG. 3). Relative abundance of dominant fungal saprotroph species based on percentage of total T-RFLP chromatogram area was calculated for one natural and one 26 y old created salt marsh in the Mississippi Sound over two seasons. For *S. alterniflora* the relative abundance of *Phaeosphaeria spartinicola* was 25% at the natural marsh and 5% at the created marsh in fall 2005 and 27% at the natural and 31% at the created marsh in spring 2006. For *Buergenerula spartinae* the relative abundance was 4% at the natural marsh and

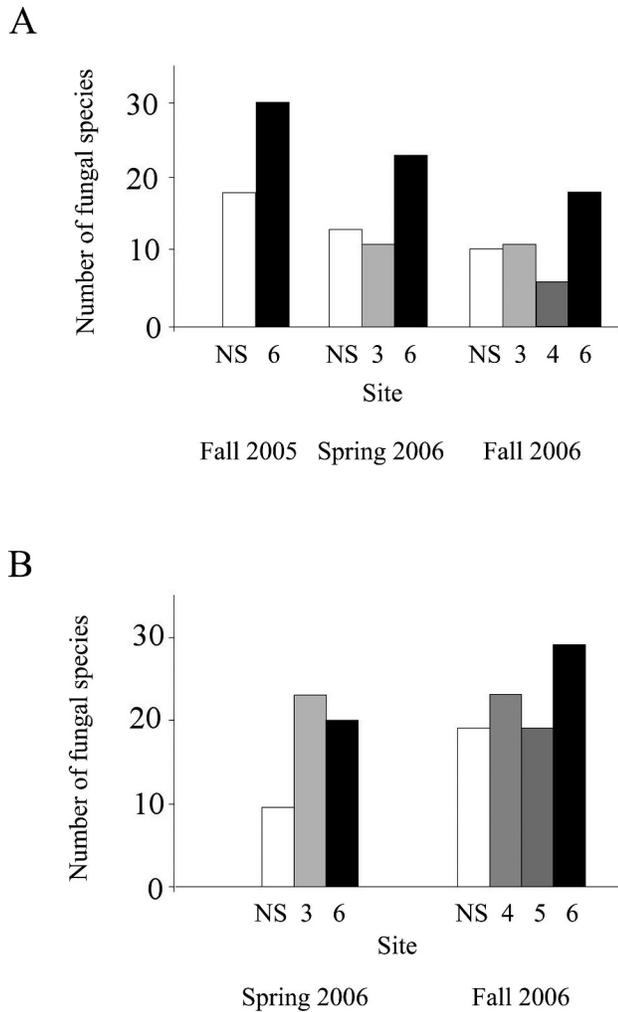


FIG. 3. A. Fungal species richness on decaying *Spartina alterniflora* at one natural salt marsh (NS) and three created salt marshes of age 1 (site 3), 1.5 (site 4) and 26 (site 6) y at the time of collection as determined by ITS T-RFLP analysis. B. Fungal species richness on decaying *Juncus roemerianus* at natural salt marshes (NS) ( $n = 2$ ) and a 1 (site 3), 1.5 (site 4), 3 (site 5) and a 26 (site 6) y old created salt marsh as determined by ITS T-RFLP analysis. Significant T-RFs were used as proxies for individual fungal species.

10% at the created marsh in fall 2005 and 6% at both the natural and the created marsh in spring 2006. Percentages are means of three samples.

#### DISCUSSION

Marine fungal saprotroph communities were present in created north-central Gulf Coast salt marshes as early as 1 y after planting. The fungal communities detected morphologically were similar to those found in East Coast salt marshes (Kohlmeyer and Volkmann-Kohlmeyer 2001, Buchan et al. 2002). The *S. alterniflora* fungal decay community of several East Coast salt marshes is dominated by *P. spartinicola* and *Mycosphaerella* sp. II (Buchan et al. 2002), while *Massarina ricifera* is a common saprotroph of *J. roemerianus* (Kohlmeyer et al. 1995). In the present Gulf Coast study *Phaeosphaeria spartinicola* and *Buergenerula spartinae* dominated the *S. alterniflora* decay community, while *Massarina ricifera* dominated the *J. roemerianus* decay system. *Mycosphaerella* sp. I on decaying *S. alterniflora* and *Phaeosphaeria roemeriani* on decaying *J. roemerianus* were collected only from natural salt marshes, suggesting that even after 26 y a created marsh has not established the same fungal community as a natural marsh. *Mycosphaerella* sp. II collected in this study had an ITS T-RF size of 15 bp, while in collections from Sapelo Island, Georgia, it had ITS T-RF sizes of 144, 410 and 428 bp (Buchan et al. 2002), suggesting that Gulf Coast and East Coast specimens might be different species and there might be multiple species within the East Coast specimens of *Mycosphaerella* sp. II.

Species richness was the first metric used to compare natural and created salt marshes. Molecular analysis revealed greater fungal species richness on both plants than was detected by morphological methods (FIG. 3A, B). The 26 y old created salt marsh at Davis Bayou Island, Mississippi, (site 6) consistently had the greatest species richness, which might indicate its fungal community contains asexual

TABLE II. Fungal isolates obtained in this study and their ITS terminal restriction fragment (T-RF) lengths

Species	Isolate	Host plant	Origin	T-RF length (bp)
<i>Mycosphaerella</i> sp. II	AK 110	<i>Spartina alterniflora</i>	MS, USA	15
<i>Phaeosphaeria spartinicola</i>	AK 109	<i>Spartina alterniflora</i>	MS Sound, USA	71
<i>Aspergillus ustus</i>	AK 117	<i>Spartina alterniflora</i>	MS Sound, USA	89
<i>Fusarium incarnatum</i>	AK 126	<i>Spartina alterniflora</i>	MS Sound, USA	148
<i>Buergenerula spartinae</i>	AK 112	<i>Spartina alterniflora</i>	MS Sound, USA	150
<i>Gloeotinia temulenta</i>	AK 120	<i>Juncus roemerianus</i>	Mon Luis, AL, USA	179
<i>Mycosphaerella</i> sp.	AK 116	<i>Juncus roemerianus</i>	MS Sound, USA	513
<i>Leptosphaeria pelagica</i>	AK 125	<i>Spartina alterniflora</i>	MS Sound, USA	631
<i>Phaeosphaeria halima</i>	AK 111	<i>Spartina alterniflora</i>	MS Sound, USA	656

TABLE III. Fungal ITS terminal restriction fragment (T-RF) lengths (in bp) detected from *Spartina alterniflora* and *Juncus roemerianus*. T-RF lengths found on both host plants are shown in boldface

Host plant	37	52	57	60	65	68	71	86	89	106	113	134	142	148	150	165	199	206	380	420	427	434	441	486	542
<i>Spartina alterniflora</i>																									
<i>Juncus roemerianus</i>																									

ascomycetes or yeasts that were not inventoried morphologically in this study, or additional sexual marine ascomycetes that were not fruiting at the time of collection. No significant effects of season, salinity, pH or dissolved oxygen on species richness were observed for either natural or created marshes.

Forty-four species of marine fungi have been detected morphologically on *J. roemerianus* in long-term studies (Kohlmeyer and Volkmann-Kohlmeyer 2001), while 33 species have been detected on *S. alterniflora* (e.g. Kohlmeyer and Kohlmeyer 1979, Leuchtman and Newell 1991). In this study five species were detected morphologically on *J. roemerianus* and six on *S. alterniflora*. However the T-RFLP community profiles generated revealed greater species richness (23 species on each host). This agrees with the pattern observed by Buchan et al. (2002), who also detected greater diversity with the T-RFLP technique at Sapelo Island, Georgia, than from morphological methods alone. Buchan et al. (2003) found that *P. spartinicola*, *P. halima*, *Mycosphaerella* sp. II and "4clt" (an unidentified species not found in the current study) together accounted for 88% of the total T-RFLP chromatogram area in more than 70% of samples (Dean Creek, Georgia) and suggested that this native fungal decay community, dominated by only a few species, can out-compete other fungal colonizers due to its adaptation to the standing decay system. A similar mechanism might be at work in Gulf Coast salt marshes.

Morphological studies have found distinct saprotrophic fungal communities present on *S. alterniflora* and *J. roemerianus* (Newell and Porter 2000). In this study six species were detected morphologically on *S. alterniflora* and five on *J. roemerianus*, none of which was detected morphologically on both plant species. However T-RFLP community profiles revealed that more than 50% of the fungal T-RFs were found on both *S. alterniflora* and *J. roemerianus*, suggesting an overlap in fungal communities on the two host plants (TABLE III). *Spartina alterniflora* and *J. roemerianus* each yielded 23 fungal T-RFs, of which 13 fragment sizes (assumed to represent separate fungal species) were common to both plant hosts. Molecular detection of the fungi indicates their presence in both host plants examined, but the absence of ascospores of the same species on both *S. alterniflora* and *J. roemerianus* suggests that certain marine fungi might be able to colonize but not sporulate on both hosts and thus might be host specific to complete their lifecycle.

Some caveats concerning T-RFLP analysis must be noted. In cases where more than one peak was observed for the same T-RF the first peak was used to obtain a representative T-RF size because the second peak was most likely the product of an incomplete

restriction digest (Clement et al. 1998). If a fungal species lacks the restriction enzyme cleavage site in the ITS gene cleavage will not occur and restriction fragments will not be generated. This could account for T-RF sizes of 651 and 656 bp (TABLE II). One solution is to use multiple restriction enzymes, each with a different restriction site. However this generates multiple community profiles, which take more time to analyze. Clement et al. (1998) found that *Hae*III yielded the largest number of TR-Fs when compared with three other restriction enzymes, making it the restriction enzyme of choice for this study. T-RFLP analysis can detect the presence of a species but not the absence of one (Dickie et al. 2002). In addition in some instances multiple species can have the same T-RF length; therefore T-RFLP community profiles can give only conservative estimates of species richness. Correlating T-RF size with species by morphological identification as done in this study helps to avoid biases introduced by sampling methods, fungal genomic structure and restriction enzyme characteristics (Avis et al. 2006).

Ascomata density of the principle marine fungus on each host was the second metric used to compare natural and created salt marshes. The presence of fungal ascomata is the end result of the decay process and indicates a large quantity of fungal biomass (hyphae) present within the plant (Newell and Porter 2000). The greater ascomata densities found in natural salt marshes during all three sampling seasons could indicate faster decay, which could in turn result in higher nutrient cycling within the ecosystem. Fungal productivity is nitrogen-limited in decaying *S. alterniflora* (Newell and Porter 2000) and created salt marshes are often nutrient poor when compared with natural salt marshes (Zedler and Lindig-Cisneros 2000), which could account for the lower ascomata densities recorded at created sites. However nutrient contents were not assessed in this study.

In conclusion saprotrophic marine fungal communities will establish rapidly in created salt marsh ecosystems but may not mirror natural communities in their composition. A combination of morphological and molecular techniques detected similar marine ascomycete communities at the natural and created north-central Gulf Coast salt marshes studied as early as 1 y after planting. ITS T-RFLP analysis revealed higher fungal species richness on both plants than morphology alone. However the sexual marine ascomycetes inventoried morphologically in this study comprise only some of the T-RFs generated; the T-RFLP profiles also may contain asexual ascomycetes found in the marine environment. *Spartina alterniflora* and *J. roemerianus* may share more than 50% of marine fungal saprotroph species, and both

were dominated by a small number of frequently occurring species. Both morphological and molecular approaches were needed to detect these patterns. Using the results of this study, saprotrophic fungal communities in additional created marshes can be examined to determine whether the marsh has an established decay community. Comparing fungal community profiles between natural and created salt marshes thus provides a qualitative measure of marsh functionality by confirming the presence of the principle decomposers of standing salt marsh vegetation.

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